



Nitrogen and Phosphorous Content in Blue Mussels (*Mytilus* spp.) Across the Baltic Sea

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OPEN ACCESS

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Specialty section:

This article was submitted to
Marine Fisheries, Aquaculture
and Living Resources,
a section of the journal
Frontiers in Marine Science

Received: 28 March 2020

Accepted: 03 August 2020

Published: 21 August 2020

Citation:

Buer A-L, Taylor D, Bergström P,
Ritzenhofen L and Klemmstein A
(2020) Nitrogen and Phosphorous
Content in Blue Mussels (*Mytilus*
spp.) Across the Baltic Sea.
Front. Mar. Sci. 7:705.
doi: 10.3389/fmars.2020.00705

To support the ongoing discussion about mussel farming and the potential to extract nutrients from the sea, this study investigated the phosphorus (P) and nitrogen (N) content of blue mussels (*Mytilus* spp.) under different abiotic and biotic parameters. The focus of this survey was on the highly eutrophied Baltic Sea, where salinity ranges from 4 to 27 psu, and is a major contributing factor to differential mussel growth. We observed that nutrient content was not linearly correlated to salinity, but if categorized, decreased at higher salinities. Chlorophyll-*a* and temperature did not significantly correlate with nutrient content, but season of harvest and mussel size did. Furthermore, habitat was a strong driver of nutrient content, indicating higher nutrient density if mussels are grown in mussel farms (i.e., in the water column) instead of on mussel culture beds or harvested from wild beds (on the sea bed). Values of N and P averaged 5.85% N and 0.83% P of tissue dry weight in mussels at the sea bed and 9.43% N and 0.96% P of tissue dry weight in mussels from longline cultivation. These results will be useful in refining estimations about mussel farming as a nutrient mitigation measure and the extraction potential, as well as related costs.

Keywords: *Mytilus* spp., Baltic Sea, nitrogen, phosphorus, salinity

INTRODUCTION

The persistence and magnitude of eutrophication in the Baltic Sea requires cost-effective measures to reduce nutrients and to achieve a good ecological status (GES) based on the Water Framework Directive (WFD, European Parliament, 2000) and the Marine Strategy Framework Directive (MSFD, European Parliament, 2008). Extensive mussel aquaculture on longlines or tube-net systems (e.g., Smart Farm) is highly discussed as such a measure in the greater Baltic Sea (Lindahl and Kollberg, 2008; Stadmark and Conley, 2011; Petersen et al., 2012, 2014, 2019; Nielsen et al., 2016; Hedberg et al., 2018; Gren, 2019; Taylor et al., 2019; Kotta et al., 2020). The amount of nutrients that can hereby be removed depends on several parameters, such as nutrient content of the mussels, growth rates, harvesting time, and farm set up (Capillo et al., 2018; Taylor et al., 2019). These parameters drive the total mitigation potential and economic feasibility of mussel farming as a eutrophication measure in different areas of the Baltic Sea. However, to the extent which environmental parameters (salinity, chl-*a*, temperature) influence the nutrient content within mussels has not been exhaustively investigated. Previous studies (Hedberg et al., 2018;

Buer et al., 2020; Holbach et al., 2020; Kotta et al., 2020) report a fluctuating nutrient content of blue mussels across the Baltic Sea but base their estimation of total mitigation potentials rather on different growth rates and an average nutrient content. Besides area-specific growth rates, it is important to evaluate the parameters that affect the actual nitrogen and phosphorus content stored in mussel tissue. Area-specific nutrient contents can support further studies on the mitigation potential, cost analyses and projections for Baltic-wide nutrient reduction.

The main objective of the present study was to document the relationship between environmental conditions and nutrient content in blue mussels across the Baltic Sea. Therefore, we analyzed the nutrient content (nitrogen, N and phosphorus, P) of blue mussels across the Baltic Sea in regards to different environmental parameters (salinity, maxima of water temperature, chlorophyll-*a* levels). Furthermore, nutrient contents were related to mussel size, habitat, and season of harvest. Providing the different environmental conditions across the greater Baltic Sea, our hypothesis was that environmental parameters would correlate with condition and nutrient contents of mussels, and an analytical approach to test for associations was followed.

MATERIALS AND METHODS

Mussel samples were collected from aquaculture farms and wild stock as indicated in **Figure 1**.

Samples along the Swedish west coast were collected at six sites from natural mussel beds, taken at 0.5 m water depth, in the spring of 2014. Mussels were divided into three size classes (5–15, 20–30, and 40–50 mm) and the dry tissue of each individual ($n = 178$) was analyzed separately for N and P content. After freeze drying, a persulfate oxidation method was used for analysis

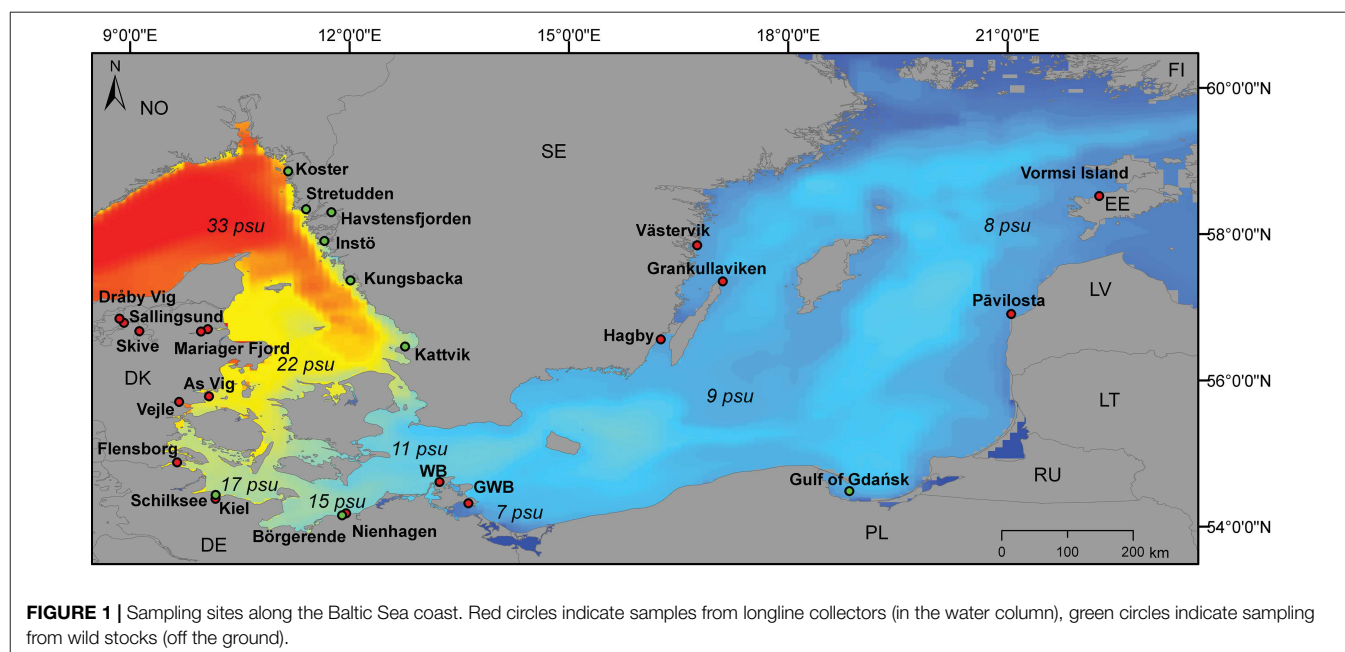
of nitrogen and phosphorous (Valderrama, 1995). The age of the mussels was estimated based on the mussel shell length.

Samples from the eastern Swedish coast (Grankullaviken, Västervik and Hagby), the Latvian coast (Pāvilosta), and Estonia (Vormsi Island) were harvested from test mussel farms within the Interreg project *Baltic Blue Growth* (BBG). Mussel samples from Greifswald Bay (GWB), Wieker Bay (WB), Nienhagen, Kiel, and Flensburg (Germany), as well as the Danish sites (Sallingsund, Skive, Dråby Vig, Mariager mid- and inner-fjord, Vejle, As Vig), were collected at mussel farms within the BONUS Optimus project. In Börgerende and Schilksee, DE, as well as Gulf of Gdańsk, PL, mussels were collected from natural hard substrate. For all sites, mussels of different size, age and season were sampled and frozen in sealed containers (-20°C) for further analyses. Subsequently, the shell length and total wet weight of a minimum of 10–30 individuals was measured. After dissection, tissue and shell of all individuals was pooled, freeze-dried, and ground, before carbon and nitrogen were analyzed in an auto-analyzer (Elementaranalysator EA 3000). Phosphorous was determined in mussel ash using an alkaline persulfate oxidation after Hansen and Koroleff (1999). Danish samples cover seven sites, while some sites were sampled in different seasons ($n = 100$ for each sampling event). The mussels were processed after ISO (1998) 6491 (1998) and ISO (2009) 5983-2 (2009) for phosphorus and nitrogen determination, respectively.

The tissue-shell-ratio, also known as the condition index, was calculated based on the dry weight of mussel tissue and shell.

$$CI = \frac{DW_{\text{tissue}} [\text{mg}]}{DW_{\text{shell}} [\text{mg}]}$$

Environmental data (salinity, temperature and chlorophyll-*a* concentration) was collected by state authority monitoring programs or modeled data (for Poland) for the upper water



layers (0.5 m depth)¹. Chlorophyll-*a* was analyzed fluorometric (665 nm) after filtration (GF/F, 0.7 μm), extraction with ethanol and acidification (ISO 10260:1992) following HELCOM – Guidelines for monitoring of chlorophyll-*a*².

Correlations of biometric data and environmental values were evaluated by bootstrapping ordinary least squares linear regression and mixed model effect leveraging by minimizing AICc. Confidence intervals were determined by bootstrapping and categorical comparison (habitat, season) was performed by permutation ANOVA (1E5 permutations) with the vegan package in R (Oksanen et al., 2019). Full factorial mixed models were bootstrapped and used to evaluate interactions with environmental variables. Normality of residuals was checked by a normal QQ plot and the Shapiro-Wilk test. Finally, for nutrient content of dry tissues, a full factorial interaction model of all environmental parameters, habitat, and season was reduced by elastic net regularization in the glmnet package in R by setting the alpha parameter to 0.5 (Friedman et al., 2010).

Additionally, data was pooled into three salinity categories: below 15 psu (6–10 psu, based on sample availability), above 15 psu and above 20 psu. Salinities below 15 psu were expected to have a strong effect on mussel condition and water content (DW/WW-ratio) (Maar et al., 2015). As a bottleneck for water exchange between the Kattegat and Baltic proper, the Great Belt area provides a natural border between salinities above and below 20 psu and thereby, as a natural divider of samples (Feistel et al., 2009; Riisgård et al., 2012). These categories were used to detect potential differences in mussel N and P content (in tissue DW and total WW) as well as differences in the tissue-shell-ratio.

RESULTS

Nutrients accumulate primarily in the bivalve tissues, only a minor fraction is stored in the shell (ratio of 0.16 ± 0.04 N in DW). Therefore, phosphorous and nitrogen concentrations of the mussel shell was not measured in all samples. To estimate N and P-values of the total mussel wet weight (including shell), measured N and P-values of the dry tissue were combined with literature data for N and P in the dry shell, averaged water contents and averaged shell-tissue-ratios. 0.04% P and 1% N of dry shell mass were assumed for all sites based on literature and our own data (Haamer, 1996; Ek Henning and Åslund, 2012; Petersen et al., 2014; Bucefalos, 2015; Palm et al., 2015; Hedberg et al., 2018). Water contents and shell-tissue-ratios were averaged for samples at each site. Results are presented in **Table 1**.

Tissue-shell-ratios were not linearly correlated to average salinity ($R^2 = 1.7e-4$, $p = 0.84$), although a weak trend toward higher ratios in higher salinities appeared if categorically pooled

(**Figure 2A**). Average site temperature was equally poor in explaining variation ($R^2 = 0.007$, $p = 0.12$) as well as chlorophyll-*a* ($R^2 = 0.01$, $p = 0.05$). Comparison of habitats exhibits a higher ratio in mussels grown in the water column [**Figure 2B**, $F_{(1,214)} = 56.61$, $p < 0.001$]. This suggests that mussels grown on suspended substrate in the water column have either thinner shells or higher somatic tissue mass compared to mussels from the sea bed. Nitrogen [$F_{(1,213)} = 610$, $p < 0.001$] and phosphorus [$F_{(1,213)} = 11.54$, $p < 0.001$] contents were similarly significantly higher in samples from mussel farms [9.55 ± 0.73 (95%CI:9.30–9.78)% N DW_{tissue}, 0.96 ± 0.17 (95%CI:0.89–1.01)% P DW_{tissue}] compared to mussels from the sea bottom [5.85 ± 0.83 (95%CI:5.72–5.96)% N DW_{tissue}, 0.82 ± 0.22 (95%CI:0.79–0.85)% P DW_{tissue}] (**Figure 2C**). In a mixed model analysis of salinity and habitat, habitat exhibited significant leverage ($R^2 = 0.50$, $p < 0.001$), but interaction between habitat and salinity was not significant ($p = 0.21$). These results suggest position in the water column or mode of attachment influences condition index to a greater degree than salinity alone. Crossing all environmental site averages was relatively effective in describing variability in the tissue-shell ratio ($R^2 = 0.38$, $p < 0.001$), where chlorophyll-*a*, temperature, and their interaction were the only significant effects ($p < 0.01$).

The nutrient content (N&P) within mussel dry tissue was weakly correlated to average salinity ($R^2_N = 0.28$, $R^2_P = 0.08$), both exhibited negative slopes, indicating a general inverse trend of nutrient content to salinity. After samples were pooled into salinity categories, the phosphorus content showed significant differences between the lowest (6–10 psu) and the intermediate (15–20 psu) as well as the highest (20–30 psu) salinity ($p < 0.001$) but not between the latter ($p = 1.000$). The nitrogen content followed the same pattern, and showed significant differences between the lowest salinity and the two higher ($p < 0.001$) but not between the intermediate and high salinity categories ($p = 0.179$) (**Figure 3A**). Differences in nitrogen in total wet weight were similar, but not significant ($p > 0.05$). Phosphorus concentrations in total wet weight were not significantly different based on salinity ($p > 0.05$) (**Figure 3B**). No correlation was observed between nutrient content in tissue dry weight to mean values of chlorophyll-*a* concentration (**Figure 3C**) nor to maximal water temperatures (**Figure 3D**) at each site. However, differences were observed between seasons (**Figure 3E**, samples of all study sites included), indicating a significant lower nitrogen content in spring compared to all other seasons [$F_{(3,211)} = 80.76$, $p < 0.001$]. Phosphorus contents on the other hand were not different between seasons [$F_{(3,211)} = 1.805$, $p = 0.1418$] (**Figure 3E**). Furthermore, nutrient content also differed between mussel sizes (**Figure 3F**, pooled samples of Swedish sites). Small mussels (5–15 mm) had a significant higher nitrogen content in the dry tissue compared to medium (20–30 mm, $p = 0.009$) and large (20–50 mm, $p < 0.001$) mussels. On the contrary, the phosphorus content increased with mussel size ($p > 0.001$ between each group).

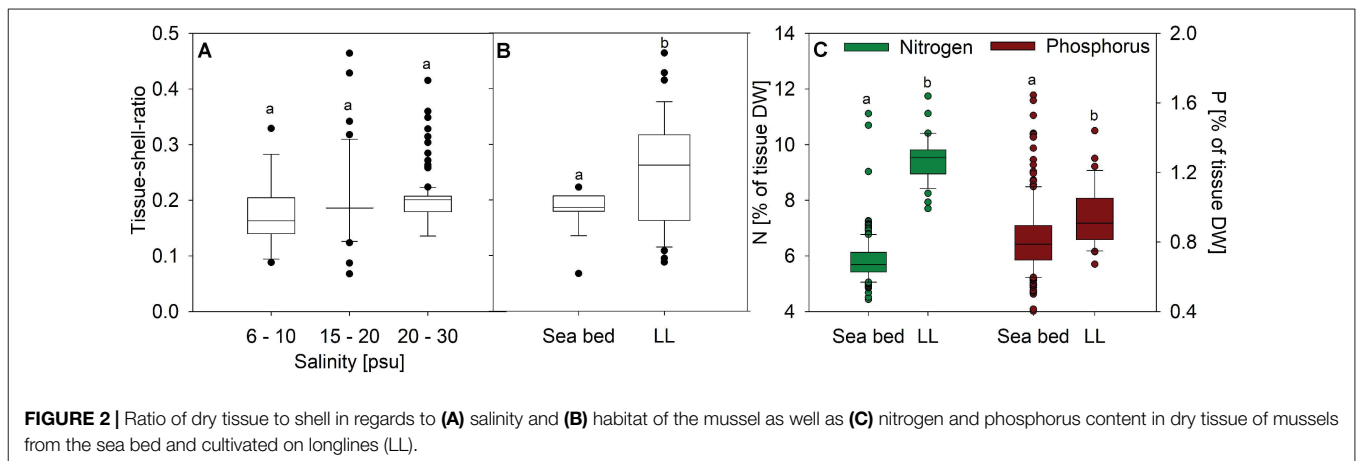
Reduction of the full factorial model for N in DW tissue, with mean environmental parameters, habitat, and season, included season*habitat, season*temperature*chlorophyll-*a*, season*salinity, and habitat as significant effects ($R^2 = 0.83$,

¹Swedish Meteorological and Hydrological Institute (SMHI, www.smhi.se/data); Latvian Institute of Aquatic Ecology; Estonian Marine Institute; Dansih Overfladevandsdatabasen (ODA, <https://oda.dk>); SatBaltyk IOPAN (satbaltyk.iopan.gda.pl); Landesamt für Landwirtschaft, Umwelt und ländliche Räume des Landes Schleswig-Holstein (LLUR); Landesamt für Umwelt, Naturschutz und Geologie Mecklenburg-Vorpommern (LUNG).

²<https://helcom.fi/media/publications/Guidelines-for-measuring-chlorophyll-a.pdf>

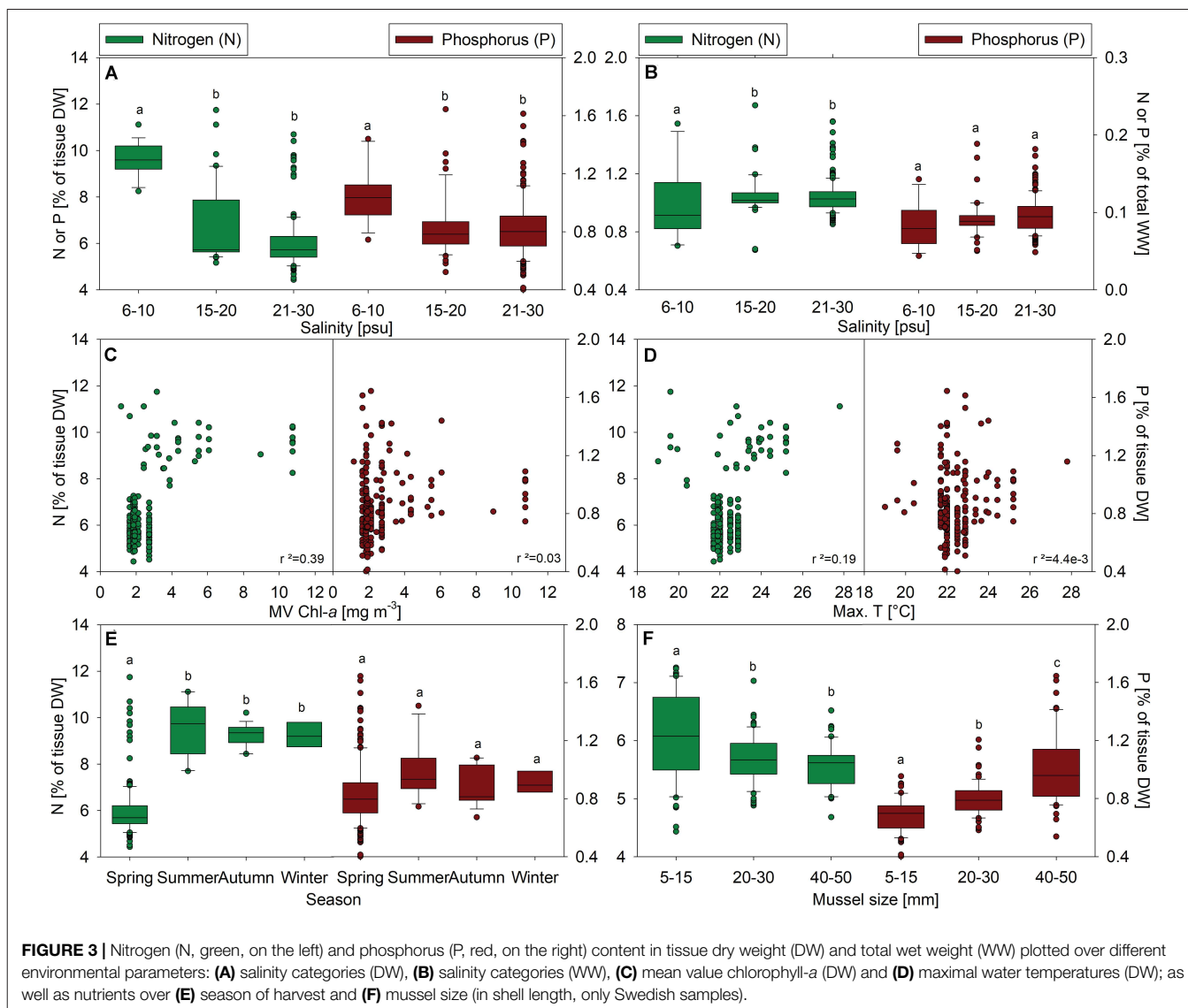
TABLE 1 | Sample locations, bivalve habitat (LL, longline; SF, Smartfarm), environmental parameters, and N&P contents in dry tissue mass (DW) as well as total wet weight (WW).

Site	Lat	Long	Habitat	Temp	Chl-a	Sal [psu]	N	P	N	P
	(y)	(x)		(°C)	(mg/m ³)	Mean	(%DW _{tissue})	(%WW _{Total})		
As Vig, DK	55.7792	10.0781	LL, SF	2.2–19.9	0.3–63	22.7 ± 2.7	9.3	0.8	1.3	0.09
Börgerende, DE	54.1539	11.8983	Pier	−0.3–22.8	0.1–16.5	15.4 ± 4.8	8.6	1.0	0.9	0.07
Dråby Vig, DK	56.8425	8.8524	LL, SF	2.0–23.9	0.3–40	27.4 ± 1.2	9.7	0.9	1.3	0.10
Flensburg, DE	54.8774	9.6408	LL, SF	0.5–23.3	0.3–73.7	17.8 ± 2.0	8.4	0.7	1.2	0.09
Grankullaviken, SE	57.3530	17.1050	LL	−0.3–22	0.1–8.8	6.9 ± 0.5	9.9	1.1	1.5	0.13
Greifswald Bay, DE	54.3172	13.6270	LL	−0.4–25.2	1.0–54.3	7.3 ± 0.7	9.5	1.0	0.8	0.06
Gulf of Gdańsk, PL	54.4853	18.8385	Sea bed	−0.4–23.6	0.2–18.1	7.3 ± 0.3	9.0	1.4	0.9	0.10
Hagby, SE	56.5600	16.2580	LL	−0.1–22.3	0.7–20	6.8 ± 0.4	8.4	1.1	1.1	0.09
Havstensfjorden, SE	58.2980	11.7519	Sea bed	0.3–22.9	0.1–31.2	26.2 ± 4.8	5.9	0.8	1.1	0.10
Instö, SE	57.9054	11.6571	Sea bed	−0.5–22	0.1–26.5	22.6 ± 5.0	5.6	0.9	1.0	0.10
Kattvik, SE	56.4607	12.7606	Sea bed	−1.1–22	0.1–16.4	19.7 ± 4.8	5.8	0.8	1.0	0.09
Kiel, DE	54.3797	10.1683	LL	0.1–19	0–38.6	17.0 ± 2.5	8.7	0.8	1.1	0.09
Koster, SE	58.8569	11.1614	Sea bed	−0.8–22.5	0.1–30	29.5 ± 4.0	6.0	0.7	0.9	0.07
Kungsbacka, SE	57.3645	12.0130	Sea bed	−0.2–21.7	0.1–36.9	26.3 ± 6.2	5.7	0.9	1.1	0.11
Mariager inner-fjord, DK	56.6664	9.9681	LL, SF	−0.5–21.9	0.3–130	15.4 ± 0.9	9.0	0.8	1.4	0.11
Mariager mid-fjord, DK	56.6990	10.0618	LL, SF	−0.9–19.6	0.7–15	18.4 ± 1.9	10.3	1.1	1.4	0.13
Nienhagen, DE	54.1798	11.9536	Sea bed	−0.3–22.8	0.1–16.5	15.4 ± 4.8	11.1	0.9	1.0	0.06
Nienhagen, DE	54.1798	11.9536	LL	−0.3–22.8	0.1–16.5	15.4 ± 4.8	8.5	0.7	1.1	0.08
Pāvilosta, LV	56.9066	21.0512	LL	−0.4–23.4	0.5–7.6	6.9 ± 0.3	9.4	1.0	1.5	0.14
Sallingsund, DK	56.7849	8.9153	LL, SF	−0.8–23.4	0.3–40	28.7 ± 1.5	9.5	0.8	1.2	0.09
Schilksee, DE	54.4329	10.1692	Pier	−0.2–20.4	0–28.1	17.8 ± 3.0	7.8	0.9	0.7	0.07
Skive, DK	56.6706	9.1262	LL, SF	0.1–24.4	0.3–65	25.1 ± 1.9	9.6	0.9	1.2	0.10
Stretudden, SE	58.3382	11.4067	Sea bed	−0.7–21.9	0.1–26.4	29.2 ± 4.3	6.0	0.7	1.1	0.09
Västervik, SE	57.8450	16.7570	LL	0.0–22.5	0.7–13	6.8 ± 0.4	10.4	1.2	1.4	0.13
Vejle, DK	55.7044	9.6687	LL, SF	−0.9–23.7	0.4–51	22.6 ± 3.4	8.9	0.8	1.4	0.10
Vormsi Island, EE	58.5190	22.2550	LL	1.4–27.8	0.1–7.6	6.7 ± 0.5	11.1	1.2	1.2	0.09
Wieker Bay, DE	54.6102	13.2322	LL	0.3–24.0	0.7–25.2	9.0 ± 0.8	9.7	1.1	0.9	0.08



$p < 0.05$). For P in DW tissue, only habitat ($p = 0.003$) and season ($p = 0.006$) had significant effects, but were weakly correlated with the data ($R^2 = 0.15$). It should be noted that habitat is correlated with chlorophyll-*a*, as longline (LL) exhibited higher mean values [5.80 ± 3.06 (95%CI:4.79–6.77)] than sea bed sites [2.05 ± 0.39 (95%CI:1.99–2.11)] [$F_{(1,213)} = 253.2$, $p < 0.001$]. Furthermore, LL sites had higher mean temperatures

[10.45 ± 0.99 (95%CI:10.14–10.78)] than sea bed sites [9.91 ± 0.31 (95%CI:9.87–9.96)] [$F_{(1,213)} = 35.45$, $p = 0.003$] and lower salinity [15.59 ± 8.55 (95%CI:12.87–18.29)] than sea bed sites [25.07 ± 3.65 (95%CI:24.55–25.61)] [$F_{(1,213)} = 116.6$, $p < 0.001$]. In sum, season and habitat were the most influential effects on the variation in N and P in DW tissue. When limiting to LL, as will be the habitat of mitigation culture under current recommendations, none of the



environmental parameters were correlated with N or P content in DW tissue, with the exception of weak negative correlation of salinity and P content ($R^2 = 0.1$, $p = 0.04$). Season as a mixed effect did not contribute to variation.

DISCUSSION AND CONCLUSION

Generally, the nutrient content within mussel tissue was not strongly related to the environmental parameters: salinity, average chlorophyll-*a* levels, or maximum water temperatures. A lower dry weight nutrient content in mussels of the eastern Baltic Sea by decreasing salinities was not observed. Consequently, our original hypothesis that these environmental parameters would account for most of the variability in mussel nutrient content, was rejected. On the other hand, it was evident that mussels cultivated on suspended substrate in the upper water column exhibited considerably higher nutrient contents than

mussels on the sea bed. It has previously been demonstrated that cultivation mode, or in general terms, position in the water column, significantly influences the biochemical composition of mussel tissues (Colombo et al., 2016). Interactions of parameters suggested that simply season and habitat encompassed much of the variability in N and P content; while it is acknowledged that some of the associated variability within the environmental parameters are built into habitat and season. Vertical gradients in phytoplankton concentrations and stratification can exacerbate food limitation in mussels at the sea bed (Wiles et al., 2006; Filgueira et al., 2018). Prior work has demonstrated a relationship between food concentration, composition, and mussel tissue biochemistry (Pleissner et al., 2012). Two potentially important implications of this habitat differentiation are: (1) samples from wild mussel beds do not serve as reliable surrogates of nutrient content when estimating potential nutrient yields in mitigation culture; (2) nutrient content of DW tissue does not appear to vary much in response to the salinity gradient.

Harvest timing (seasonality) is an important parameter when considering the potential nutrient extraction potential. Low spring nitrogen contents in cultivated blue mussels are consistent with primary gamete release (Pieters et al., 1980; Okumus and Stirling, 1998; Fernández et al., 2015). Taylor et al. (2019) recommended late autumn through winter harvests in order to maximize nutrient extraction efficiency. This will additionally avoid potential loss of nutrient content overlapping with gamete release, and permit new spat collection in the late spring/early summer.

Due to the low resolution of environmental parameters in time and space, it was difficult to draw definitive conclusions from observed relationships of nutrient content and environmental conditions. These analyses could be improved by evaluating interannual variation relative to position-specific high resolution environmental data. Supplementing interpretations of nutrient extraction potentials by energetic models may corroborate the importance of environmental conditions to nutrient content (Buer et al., 2020).

In sum, these results suggest that variability in proportional mussel nutrient content will be nominal when cultivated in the water column if there is sufficient food available, and harvesting timed to avoid the reproduction period. Total biomass, therefore, will be the meaningful metric when estimating nutrient extraction potential. Furthermore, although there is reduced growth of blue mussels at low salinities, proportional nutrient content is not significantly reduced in the summary harvested dry materials.

DATA AVAILABILITY STATEMENT

All datasets generated for this study are included in the article/**Supplementary Material**.

AUTHOR CONTRIBUTIONS

A-LB was responsible for overall structure and writing of the manuscript, data collection, and sample processing. DT and LR

contributed to data curation and manuscript writing. PB and DT provided the data for nutrients in mussels of Sweden's west coast and Denmark, respectively, who also contributed to the text of the manuscript. AK was involved in sample processing. All authors contributed to the article and approved the submitted version.

FUNDING

The work was financially supported by the project BONUS OPTIMUS (03A0020A). The project has received funding from BONUS (Art 185) funded jointly by the European Union's Seventh Programme for research, technological development and demonstration, and from Baltic Sea national funding institutions. The publication of this article was funded by the Open Access Fund of the Leibniz Association.

ACKNOWLEDGMENTS

We would like to thank the people who provided mussel samples from around the Baltic: Juris Aigars (Latvian Institute of Aquatic Ecology), Jonne Kotta (University of Tartu), Susanna Minnhagen (Kalmar Municipality), Ksenia Pazdro (IOPAN), Tim Staufenberger (Kieler Meeresfarm), and Florian Peine (LFA). For the provision of environmental long-term data, we thank SMHI, Latvian Institute of Aquatic Ecology, Estonian Marine Institute, Danish Overfladevandsdatabasen, IOPAN, LLUR, and LUNG. Furthermore, Regina Hansen, Anne Köhler, and Ines Scherff for their support on CNP analyses, as well as Ivar Lund and Ulla Sproegel for Danish nutrient analyses. Additionally, we would like to thank the editor and the three referees for their critical review and very valuable input.

SUPPLEMENTARY MATERIAL

The Supplementary Material for this article can be found online at: <https://www.frontiersin.org/articles/10.3389/fmars.2020.00705/full#supplementary-material>

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Conflict of Interest: The authors declare that the research was conducted in the absence of any commercial or financial relationships that could be construed as a potential conflict of interest.

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